
SHORT COMMUNICATION

Spatial and temporal variations in percentage cover of two common seagrasses at Sirinart National Park, Phuket; and a first step for marine base

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Abstract

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Percentage cover and distribution of two common seagrasses, *Thalassia hemprichii* (Ehrenb.) Aschers. and *Cymodocea rotundata* Ehrenb. Et Hempr. Ex Aschers., were studied in the dry and wet seasons. The study was carried out at three levels on sheltered, moderately exposed and very exposed sites on the coastline of Sirinart National Park, Thailand. One hundred and twenty samplings were investigated and recorded. Analysis of variance (ANOVA) revealed that there were significant differences in the percentage cover of *C. rotundata* among different degrees of wave exposure ($P<0.01$); and *T. hemprichii* was significantly influenced by interactions between seasons, shore levels and degrees of wave exposure ($P<0.05$). High sediment disturbance on the very exposed site was likely to influence the percentage cover and distribution of both seagrasses. This study provided baseline data for further work on ecological study and long term monitoring; and a first step to building up a 'marine base'.

Key words : *Cymodocea rotundata*, *Thalassia hemprichii*, marine ecology and marine base

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บทคัดย่อ

อัญชนา ประเทพ
ความต่างของพื้นที่ และฤดูกาลต่อปริมาณหญ้าทะเล 2 ชนิด ณ อุทยานแห่งชาติสิรินาถ
จ.ภูเก็ต และฐานข้อมูลเบื้องต้นสำหรับการสร้างศูนย์วิจัยนิเวศวิทยาทางทะเล
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การศึกษาปริมาณและการแพร่กระจายของหญ้าทะเล 2 ชนิด คือ *Thalassia hemprichii* (Ehrenb.) Aschers. และ *Cymodonea rotundata* Ehrenb. Et Hempr. Ex Aschers. ณ อุทยานแห่งชาติสิรินาถ จ.ภูเก็ต โดยทำการศึกษาผลของปัจจัยทางกายภาพต่อปริมาณและการแพร่กระจายของหญ้าทะเล โดยปัจจัยที่ศึกษาต่างกันดังนี้ 1) ฤดูกาล (ฤดูร้อน ฤดูฝน) 2) ปริมาณคลื่นลม (บริเวณอันล่ม บริเวณได้รับคลื่นลมปานกลาง และบริเวณที่ได้รับคลื่นลมมาก) และ 3) ระยะทางห่างจากฝั่ง (น้ำขึ้นสูงสุด) (0-40 ม. 41-80 ม. และ 81-120 ม.) จากการสู่นตัวอย่าง 120 ครั้ง พบว่า *C. rotundata* ได้รับอิทธิพลจากปริมาณคลื่นลมอย่างมีนัยสำคัญ ($p<0.01$) และ *T. hemprichii* ได้รับอิทธิพลร่วมจากทั้ง 3 ปัจจัยอย่างมีนัยสำคัญ ($p<0.05$) ความแตกต่างดังกล่าวอาจมีผลมาจากการปริมาณตะกอนในมวลน้ำซึ่งได้รับอิทธิพลโดยทางอ้อมจากความแรงของคลื่นลม การศึกษานี้เป็นข้อมูลพื้นฐานสำหรับการศึกษานิเวศวิทยาขั้นสูง และการติดตามผลเชิงนิเวศวิทยาระยะยาว พร้อมกันนี้เป็นการสร้างฐานข้อมูลสำหรับศูนย์วิจัยนิเวศวิทยาทางทะเล

ภาควิชาชีววิทยา คณะวิทยาศาสตร์ มหาวิทยาลัยสงขลานครินทร์ อําเภอหาดใหญ่ จังหวัดสงขลา 90112

Seagrasses are underwater flowering plants that form an important coastal habitat world wide, often occurring in vast meadows which provide nurseries, shelter, and food for a variety of commercially, recreationally, and ecologically important species. There are 12 genera, in which 48 species have been reported worldwide (Phillips and Menez, 1988), 16 species from the ASEAN region (Fortes, 1990); and 12 species in Thailand (Lewmanomont *et al.*, 1996; Santisuk and Larsen, 2001) which support many endangered species such as turtles and dugongs. However, there have been only a few studies conducted on the species composition and distribution of seagrasses in Thailand (Changsang and Poovachiranon, 1994; Poovachiranon and Changsang, 1994; Lewmanomont *et al.*, 1996; Meesawat *et al.*, 1999; Nakaoka and Supanwanid, 2000). Moreover, no long term of monitoring seagrasses have been yet reported. Sharp increases of the sediment in the water column due to coastal developments, deforestation and erosion can cause dramatic changes to seagrasses and other marine communities (e.g. Airoldi *et al.*, 1996; Duarte, 2001).

The ecological roles of seagrasses are very important, they filter estuarine and coastal waters

of nutrients, contaminants, and sediments, and are closely linked to other communities e.g. coral reef and mangrove systems (Nybakken, 2001). Moreover, they provide habitats for a wide variety of marine organisms, both plants and animals. The relatively high rate of primary production in seagrasses drives detritus-based food chains, which helps to support many of these organisms (Adam and King, 1995). Due to the important roles of seagrasses, there is 'SeagrassNet' which is a global monitoring program to investigate and document the status of seagrass resources world wide, which have been studied intensively in ASEAN water, but mostly in the Philippines (e.g.; Terrados *et al.*, 1999a; Terrados 1999b; Durate *et al.*, 2000; Agawin *et al.*, 2001; Van Vierseen *et al.*, 2001; Husan *et al.*, 2002; Kamp-Nielsen *et al.*, 2002; Lacap *et al.*, 2002). However, more studies on a local scale are still needed, which would be an important baseline dataset for Thai seagrasses.

Physical factors such as wave action and desiccation are already known to play very important roles in the distribution and abundance of most intertidal marine organisms (Lewis, 1964), but only a few such studies have been made on seagrass communities (Nakaoka and Supanwanid,

2000). Thus, this study investigated the effects of degree of wave exposure and shore level on distribution and abundance of seagrasses at the Sirinart National Park, Phuket, Thailand. This will provide essential baseline information for further complex ecological study. In addition, this is a first step to build up baseline information for further long term monitoring by using Sirinart Marine Park as a 'marine base'.

Materials and Methods

Study site and sampling

The study site was located at Koh Pling, Sirinart National Park, Phuket, South of Thailand ($8^{\circ}05'N$, $98^{\circ}17'E$). Sampling sites were selected along the shoreline with different degrees of wave exposure: sheltered (S), moderately exposed (M) and very exposed (E), at Ban Sakuu. At the exposed area, organisms were directly exposed to the wave action, which was less in mid-exposed and sheltered areas due to protection by fringing reefs. The study was carried out during low tide when most grass beds were exposed. Six line transects were conducted among the different degrees of wave exposure within the grass beds, two lines each. Then each line was marked using A+B EpoPutty epoxy (ALTECO) fixing on the rocks individually. Three or four quadrats of $50\text{ cm} \times 50\text{ cm}$ were sampled randomly at 40 m intervals at three shore levels: 0-40 m was upper level, 41-80 m was mid shore

level and 81-120 m was lower shore level. The tidal range at Phuket was 0.8-3.8 m in 2002; mean sea level was about 2.3 m (calculated from the Tide Table of the Hydrographic Department, Royal Thai Navy). Samples were monitored and recorded in two seasons: a dry season predominated by the NE Monsoon and a wet season predominated by the SW Monsoon. Dry season study was conducted during 6-8 May 2002 and wet season study was conducted during 4-6 October 2002. This brought up to 120 samples for this study. Percentage cover and substrates of *Cymodocea* and *Thalassia* were estimated visually and recorded at the site. Macrophyte specimens were collected and taken to the laboratory for species identification using Common Seaweeds and Seagrasses of Thailand (Lewmanomont and Ogawa, 1995) and Flora of Thailand (Santisuk and Larsen, 2001).

Statistical analysis

Two-way ANOVA was employed to test percentage cover of each species against different sites and seasons. Multiple comparisons were tested when there were significant differences between treatments, following Zar (1984). Cochran's C-test was used before each analysis, to test whether variances were homogeneous and square-root transformation was applied when necessary. Tukey multiple comparison was employed to test the differences between sites and seasons.

Table 1. Analysis of variance on mean percentage cover of *Thalassia hemprichii* and *Cymodocea rotundata* at Koh Pling, Sirinart National Park, Phuket.
* $P<0.05$, ** $P<0.01$, *** $P<0.001$

Source of Variation	df	<i>Thalassia hemprichii</i>		<i>Cymodocea rotundata</i>	
		MS	F	MS	F
Season: S	1	365.02	0.46	0.01	0.99
Shore level: SL	2	2161.71	0.04*	41.81	0.79
Degree of exposure: EX	2	2338.87	0.35*	933.84	0.006**
S × SL	2	589.87	0.41	68.89	0.67
S × EX	2	318.60	0.62	2.44	0.99
SL × EX	2	3781.95	0.0004***	34.88	0.94
S × SL × EX	4	2235.20	0.014*	75.71	0.78

Results

Species composition and distribution

Two species of seagrasses were found mixed together along the transect lines on coarse sand, *Thalassia hemprichii* and *Cymodocea rotundata*. *C. rotundata* was recorded mostly on the sheltered area than the other sites. While, *T. hemprichii* was widely distributed at all sites. Also, it is worth noting that there were common brown algae, *Padina* spp., found on the dead corals or rock within and nearby the grass beds.

Spatial and temporal variations in populations

There was variation in the percentage cover of *T. hemprichii* and *C. rotundata* amongst sites and seasons. *T. hemprichii* was the most dominant macrophyte at the study area, while *C. rotundata* was sparse and occurred mainly on the sheltered shore. Both seagrasses, however, had very little seasonal change, but they were significantly influenced by shore level and degree of exposure (Table 1). *T. hemprichii* was widely distributed along the shore, the highest percentage cover was 85.56% on the high shore of the mid exposed area during the dry season (Figure 1). *C. rotundata* was significantly more abundant on the sheltered shore than at the other sites, the highest percentage cover was only 14.16%. (Figure 2, Table 1).

Discussion

This study was the first study to investigate species diversity, distribution, spatial and temporal variations of seagrasses; and it was an attempt to build up baseline data for long term marine ecological study by using Sirinart National Park as a 'marine base'. Two species of seagrasses were found at the study site, *T. hemprichii* and *C. rotundata*. They are commonly found along the Andaman sea coast of Thailand (e.g., Changsang and Poovachiranon, 1994; Meesawat *et al.*, 1999; Lewmanomont and Supanwanid, 2000; Nakako and Supanwanid, 2000); and they are two of the most common species in South-East Asian seagrass meadows (Terrados *et al.*, 1999b).

Both seagrasses were very patchy, the area cover of *T. hemprichii* was 0.043 km² - 0.069 km² and *C. rotundata* was 0.017 km². The area cover of *T. hemprichii* was two times greater eight years previously, when compared to the study by Changsang and Poovachiranon (1994). However, there was no record of *C. rotundata* in 1994. Thus, *C. rotundata* was likely to have dispersed from nearby areas and have established a new community recently. To my knowledge, there are only a few studies on propagule dispersal of seagrasses (e.g., Lacap *et al.*, 2000). Therefore, further study on propagule dispersal and recruitment of seagrasses are needed for further understanding in their distribution and establishment.

Here, the seagrass bed was very small compared to the 18 km² of the seagrass bed at Ko Muk, a very big seagrass bed in Thailand; also it was much lower in species richness, with only two species compared to the seven species recorded at Ko Muk (Nakaoka and Supanwanid, 2000). However, there was still quite a variety of marine organisms within this seagrass bed; and there was much greater abundance of fauna than on the adjacent bare sediments (personal observation). This might, therefore, be a result of their productivity and the complexity of the seagrasses, which provide both food sources and shelter for the faunas (Heck and Thoman, 1981; Robertson, 1984, Nelson and Bonsdorff, 1990; Irlandi *et al.*, 1995; Boström and Mattila, 1999).

Unlike *C. rotundata*, *T. hemprichii* covered a wide range on the shore, from the sheltered shore to the very exposed shore, where there was high disturbance from sediment movement. However, the highest abundance was found on the moderately exposed shore during the dry season, at the high shore level. This might be a result of reduced disturbance from the sediment movement that caused high water turbidity and stresses to the seagrass. Shading due to high turbidity can cause decreased primary productivity of seagrass. Moreover, sediment movement influenced distribution, abundance and productivity of seagrasses (e.g. Duarte *et al.*, 1997; Vermaat *et al.*, 1997; Terrados *et al.*, 1998) and other marine organisms (see

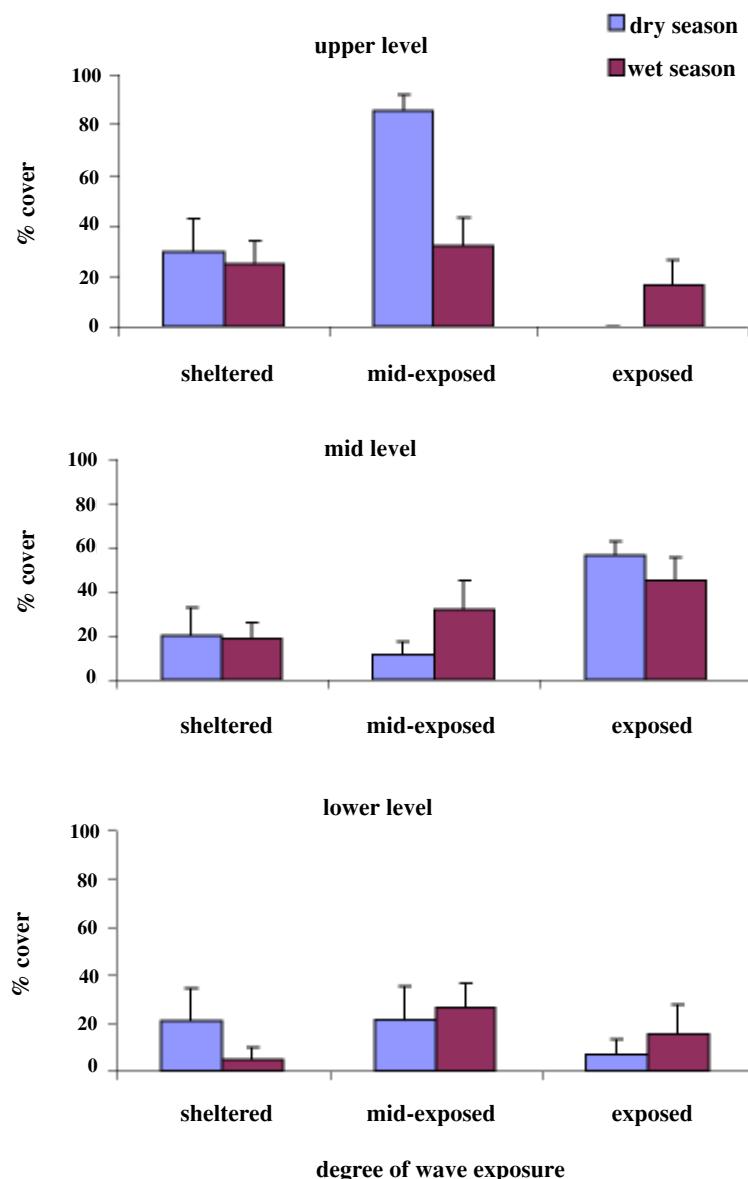


Figure 1. Effects of degrees of wave exposure and levels in two seasons on percentage cover of *Talassia hemprichii*, at Srinart National Park, Phuket.

Prathee *et al.*, 2003).

The establishment of a new bed at the sheltered area, where it was protected from the wave exposure, indicated a favorable habitat for the *C. rotundata*. However, it is unlikely to over-grow *T. hemprichii* due to the competitive superiority of *T. hemprichii* (Brouns, 1987, Vermaat *et al.*, 1995).

Two main processes are responsible for apparent seagrass loss: 1) A natural shift of the bed as part of a natural dynamic trend, producing a loss of seagrasses in one particular area yet an increase in another. A dynamic trend "loss" may appear as a net loss in one area and net gain in another but, in fact, is merely a shift of the seagrass bed resulting in no net loss or gain. 2) Loss

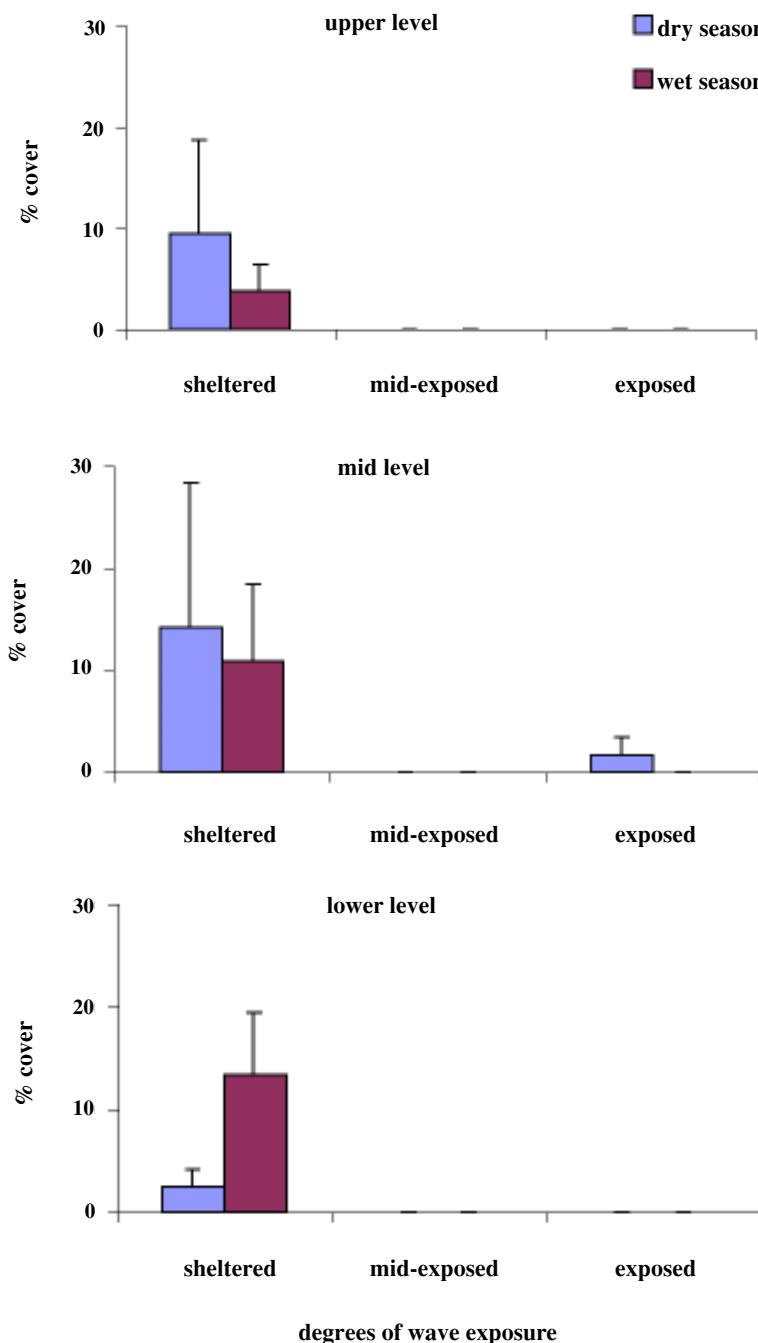


Figure 2. Effects of degrees of wave exposure and levels in two seasons on percentage cover of *Cymodocea rotundata* at Sirinart National Park, Phuket.

due to the weather and human disturbances resulting in a net loss of seagrass. This category includes both natural and anthropogenic causes (Florida

Department of Environmental Protection, 2001). Humans have the potential to greatly disrupt the seagrass ecosystem. Generally, this ecosystem

is adapted to cyclic natural phenomena such as changes in temperature, light, and nutrients. In contrast, human activities may be continuous or episodic events, to which organisms are not adapted, e.g. trawling, dredging, and nutrient inputs (Texas and Wildlife, 1999). I have also seen many local people fishing and collecting clams nearby the bed. They might directly disturb the seagrass bed by tramping on the grass or indirectly collect other marine organisms which caused decrease in diversity and an imbalance in the seagrass ecosystem. Unlike the coral reef ecosystem, seagrass and rocky shore ecosystems seem to have less appeal for others even though they are very important primary producers and shelters on coastal areas. Therefore, further research and education on the importance of seagrasses for local people on the seagrass beds and the rocky shore are much needed.

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References

Adam, P. and King, R.J. 1995. Ecology of unconsolidated shores. In *Biology of Marine Plants*. (Eds). M.N.Clayton, R.J. King. Longman Australia Pty Limited, Melbourne, pp.296-309.

Agawin, N.S.R., Duarte, C.M., Fortes, M.D., Uri, J.S. and Vermaat, J.E. 2001. Temporal changes in the abundance, leaf growth and photosynthesis of three co-occurring Philippine seagrasses. *J. Exp. Mar. Biol. Ecol.* 260: 217-239.

Airoldi L., Fabiano, M. and Cinelli, F. 1996. Sediment deposition and movement over a turf assemblage in shallow rocky coastal area of the Ligurian Sea. *Mar. Ecol. Prog. Ser.*, 133: 241-251.

Boström, C. and Mattila, J. 1999. The relative importance of food and shelter for seagrass-associated invertebrates: a latitudinal comparison of habitat choice by isopod grazers. *Oecologia*, 120: 162-170.

Brouns, J.J.W.M. 1987. Quantitative and dynamic aspects of a mixed seagrass meadow in Papua New Guinea. *Aquat. Bot.* 29: 33-47.

Changsang H. and Poovachiranon, S. 1994. The distribution and species composition of seagrass beds along the Andaman Sea coast of Thailand. *Phuket Mar. Biol. Cent. Res. Bull.* 59: 43-52.

Duarte, C.M., Terrados, J., Agawin, N.S.R., Fortes, M.D., Bach, S. and Kenworthy, W.J. 1997. Response of a mixed Philippine seagrass meadow to experimental burial. *Mar. Ecol. Prog. Ser.* 147: 285-294.

Duarte, C.M., Terrados, J., Agawin, N.S.R., Fortes, M.D. 2000. An experimental test of the presence of competitive interactions among SE Asian seagrasses. *Mar. Ecol. Prog. Ser.* 197: 231-240.

Duarte, C.M. 2001. The PREDICT project: Predicting the Recovery of SE Asian Coastal Plant Communities. *Research for Sustainable Development Bulletin*. 14(1-4): 23-26.

Florida Department of Environmental Protection. 2001. *Seagrass Management Plan for Big Lagoon and Santa Rosa Sound. Ecosystem Restoration Section*, Florida.

Fortes, M.D. 1990. Seagrasses: A Resource Unknown in the ASEAN region. ICLARM Educational Series 6, 46 p. International Center for Living Aquatic Resources Management. Manila, The Philippines.

Halun, Z., Terrados, J., Borum, J., Kamp-Nielsen, L., Duarte, C.M., Fortes, M.D. 2002. Experimental evaluation of the effects of siltation-derived changes in sediment conditions on the Philippine seagrass *Cymodocea rotundata*. *J. Exp. Mar. Biol. Ecol.* 279: 73-87.

Heck, K.L. and Thoman, T.A. 1981. Experiments on predator-prey interactions in vegetated aquatic habitat. *J. Exp. Mar. Biol. Ecol.* 53: 125-134.

Irlandi, E.A., Ambrose, W.G.Jr. and Orlando, B.A. 1995. Landscape ecology and the marine envi-

ronment: how spatial configuration of seagrass habitat influences growth and survival of the bay scallop. *Oikos* 72: 307-313.

Kamp-Nielsen, L., Vermaat, J.E., Wesseling, I., Borum, J. and Geertz-Hansen, O. 2002. Sediment properties along gradients of siltation in South East Asia. *Estuar. Coast. Shelf. Sci.* 54: 127-137.

Lacap, C.D.A., Vermaat, J.E., Rollon, R.N. and Nacorda, H.M.N. 2002. Propagule dispersal of the SE Asian seagrasses *Enhalus acoroides* and *Thalassia hemprichii*. *Mar. Ecol. Prog. Ser.* 235: 75-80.

Lewis, J.R. 1964. *The Ecology of Rocky Shores*. English Universities Press. London.

Lewmanomont, K. and Ogawa, H. 1995. Common Seaweeds and Seagrasses of Thailand. Integrated Promotion Technology Co.Ltd, Bangkok.

Lewmanomont, K., Deetae, S. and Srimanobhas, V. 1996. Seagrasses of Thailand. Proc. 2nd Seagrass Biology Workshop, Rottnest Island, Western Australia, Jan., 1996: 21-26.

Meesawat, U., Purintavarangul, C., Mayakul S., Hirunpun, R. 1999. Diversity and seasonal anatomical changes of seagrasses at Hat Chao Mai National Park, Trang Province. *Songklanakarin J. Sci. Technol.* 21: 65-81.

Nakaoka M. and Supanwanid, C. 2000. Quantitative estimation of the distribution and biomass of seagrasses at Haad Chao Mai National Park, Trang Province, Thailand. *Kasetsart University Fishery Bulletin*. 22: 1-9.

Nelson, W.G. and Bonsdorff E. 1990. Fish predation and habitat complexity; are complexity thresholds for real? *J. Exp. Mar. Biol. Ecol.* 141: 183-194.

Nybakken, J.W. 2001. *Marine Biology: An Ecological Approach*. Benjamin Cummings. San Fransico.

Phillips, R.C. and Menez, E.G. 1988. *Seagrasses*. Smithsonian Institution Press, Washington D.C., U.S.A..

Poovachiranon, S. and Changsang, H., 1994. Community structure and biomass of seagrass beds in the Andaman Sea. I. Mangrove-associated sea grass beds. *Phuket Mar. Biol. Cent. Res. Bull.* 59: 53-64.

Prathee, A., Marrs, R.H. and Norton, T.A. 2003. Spatial and temporal variations in sediment accumulation in an algal turf and their impact on associated fauna. *Mar. Biol.*, 142: 381-390.

Robertson, A.I. 1984. Trophic interations between the fish fauna and macrobenthos of an eelgrass community in Western Port, Victoria. *Aquat. Bot.* 18: 135-153.

Santisuk, T. and Larsen, K. 2001. *Flora of Thailand. Volume 7(3). The Forest Herbarium, Royal Department*. Bangkok.

Terrados, J., Agawin, N.S.R., Duarte, C.M., Fortes M.D. and Kamp-Nielsen, L. 1999a. Nutrient limitation of the tropical seagrass *Enhalus acoroides* (L.) Royle in Cape Bolinao, NW Philippines. *Aquat. Bot.* 65: 123-139.

Terrados, J., Borum, J., Duarte, C.M., Fortes, M.D., Kamp-Nielsen, L., Agawin, N.S.R. and Kenworthy, W.J. 1999b. Nutrient and mass allocation of South-East Asian seagrasses. *Aquat. Bot.* 63: 203-217.

Texas and Wildlife. 1999. *Seagrass Conservation Plan for Texas*, Resource Protection Division, Austin. TX.

Van Vierssen, W., Van Hofwegen, P.J.M. and Vermaat, J.E. 2001. The age of water scarcity: in search of a new paradigm in aquatic weed control, J. *Aquat. Plant Manage.*, 39: 3-8.

Vermaat, J.E., Fortes, M.D., Agawin, N., Duarte, C.M., Marba, N. and Uri, J., 1995. Meadow maintenance, growth and productivity in a mixed Philippine seagrass bed. *Mar. Ecol. Progr. Ser.* 124: 215-225.

Vermaat, J. E., Agawin, N.S.R., Fortes, M.D., Uri, J.S., Durate, C.M., Marba, N., Enriquez S. and Van Vierssen, W. 1997. The capacity of seagrasses to survive increased turbidity and siltation: the significance of growth form and light use. *Ambio.*, 26: 499-504.

Zar, J.H. 1984. *Biostatistical Analysis*. Prentice-Hall International, London.